

Original article

Diet and body shape among populations of *Bryconamericus iheringii* (Otophysi: Characidae) across the Campos Sulinos ecosystem

Érika Endo Kokubun¹, Karine Orlandi Bonato¹,
Edward D. Burress² and Clarice Bernhardt Fialho¹

Alterations in natural landscapes, mainly caused by anthropic pressures, have been threatening the world's biomes, including aquatic environments and its biota. This study describes the diet of *Bryconamericus iheringii*, and how its body shape relates to environmental variables in populations of 22 streams. A wide array of food items were found, mainly composed of allochthonous plants (50.5%) and autochthonous invertebrates (25.2%). Even though food items remained almost the same, the predominant food group significantly differed among streams, mainly in relation to environmental characteristics. There was variation in body shape primarily associated with body depth and length of the pre-dorsal region; however, these differences did not correspond with streams. PLS-CA analyses indicated that environmental characteristics, such as substrate type, percentage of marginal vegetation have some influence over food items availability but not on body shape. This may be because *B. iheringii* is a non-specialist species capable of prey switching based on availability due to an intermediate body shape suited for generalist feeding habits.

Keywords: Ecomorphology, Generalist diet, Grassland ecosystem, Stream.

Alterações em paisagens naturais, principalmente as causadas por pressão antrópica, tem ameaçado os biomas mundiais, incluindo ambientes aquáticos e sua biota. Este estudo descreve a dieta de *B. iheringii* e como seu formato corporal se relaciona com variáveis ambientais, em populações de 22 riachos. Uma grande variedade de itens alimentares foi encontrada, principalmente compostos por plantas alóctones (50,5%) e invertebrados autóctones (25,2%). Ainda que os itens alimentares tenham permanecido quase os mesmos, o grupo alimentar predominante diferiu significativamente entre riachos, principalmente devido a características ambientais. Ocorreu variação no formato corporal principalmente associado com profundidade do corpo e comprimento da região pré-dorsal; entretanto, tais diferenças não corresponderam aos agrupamentos por riacho. A análise PLS-CA indicou que as características ambientais, tais como tipo de substrato, porcentagem de vegetação marginal exercem alguma influência sobre a disponibilidade dos itens alimentares, mas não sobre o formato corporal. Isso pode ocorrer por *B. iheringii* ser uma espécie não especialista com capacidade de mudar suas presas de acordo a com disponibilidade das mesmas, isto, devido ao seu formato corporal intermediário adequado a hábitos generalistas.

Palavras-chave: Dieta generalista, Ecomorfologia, Ecossistema de pastagem, Riachos.

Introduction

Due to great anthropogenic pressure, the world's natural landscapes have been transformed (DeFries *et al.*, 2004; Foley *et al.*, 2005). Land use, for instance, enhances primary production available for human populations but decreases availability for ecosystem processes (DeFries *et al.*, 2004), changes hydrological responses (Costa *et al.*, 2003), causes habitat loss and fragmentation, degradation of soil, and changes the assemblage composition of plants, fish, and

birds (Harding *et al.*, 1998; Grau *et al.*, 2003; Waltert *et al.*, 2004). Given that aquatic and terrestrial environments are interdependent (Dufech *et al.*, 2003), activities that affect land cover and its ecological processes may propagate impacts to aquatic environments and their biotic components, including native ichthyofauna (Ferreira, Casatti, 2006). Alterations in physical and chemical properties of terrestrial and aquatic habitats, therefore, can modify characteristics of fish assemblages such as abundance and diversity (Tramer, Rogers, 1973; Dyer *et al.*, 2003; Cunico *et al.*, 2006).

¹Laboratório de Ictiologia, Instituto de Biociências, Universidade Federal do Rio Grande do Sul, Av. Bento Gonçalves, 9500, Agronomia, 91501-970 Porto Alegre, RS, Brazil. (EEK) erikakokubun@gmail.com, (KOB) kakabio2005@yahoo.com.br, <http://orcid.org/0000-0002-3741-8086> (corresponding author), (CBF) cbfialho@via-rs.net

²Department of Biological Sciences, Auburn University Museum of Natural History, Auburn University, 101 Life Sciences Bldg., Auburn University, 36849 Auburn, AL, USA. edwarddburress@gmail.com

Despite being an endangered biome due to agriculture and silviculture activities, the Campos Sulinos ecosystem is still considered an area with high fish diversity (Pillar, Lange, 2015). Its ichthyofauna extends throughout four drainage systems: the Uruguai River basin, Laguna dos Patos System, part of Paraná State drained by Iguaçú River, and a coastal basin between the states of Rio Grande do Sul and Santa Catarina (MMA, 2000). Yet, our knowledge of its fish fauna is fragmented and scarce in many aspects (*i.e.*, taxonomy, biogeography, ecology, and conservation status). Biological traits are useful tools to increase knowledge and to help understand certain ecological aspects. Feeding biology and morphology, for instance, are good indicators of population dynamics, ecology of species, community structure and help to understand the food web of the system (Polis, Winemiller, 1996).

Even though there can be a certain specificity for certain food items, the majority of fish species have a highly flexible feeding behavior (*i.e.* trophic plasticity) (Abelha *et al.*, 2001; Meurer, Zaniboni-Filho, 2012), particularly Neotropical freshwater species. This feeding flexibility is an adaptive response (Brandão-Gonçalves *et al.*, 2009) to fluctuations in the availability of food resources in the environment (Abelha *et al.*, 2001). Such fluctuations may be caused by several factors such as physicochemical, spatial and temporal components (Lawlor, 1980; Costa *et al.*, 2013; Jacobson *et al.*, 2015). For example, the presence of riparian vegetation is an important source of food subsidies for fishes (Esteves, Aranha, 1999; Burnett, Reeves, 2006) as well as for shading the water (Pusey, Arthington, 2003). Its absence, therefore, may result in higher levels of insolation and, consequently, increase primary production by facilitating the growth of diatoms and filamentous algae (Allan, 2004). Other possible consequences include a decrease in invertebrate density (Vasconcelos, Melo, 2008). All those alterations, nearby and within the stream, may influence the diet of fishes by favoring opportunist species that are more flexible in their foraging behaviors and preferences (Oliveira, Bennemann, 2005; Bonato *et al.*, 2012).

Morphological characteristics such as general body shape, head length and caudal peduncle aspect ratio may also be influenced by habitat alterations (Langerhans *et al.*, 2003; Neves, Monteiro, 2003; Zúñiga-Vega *et al.*, 2007), given that differences in body shape are often associated with biotic and abiotic habitat characteristics (Eklöv, Svanbäck, 2006; Costa, Cataudella, 2007; Faradonbe *et al.*, 2015). Morphology may reflect different ecological and behavioral patterns, including foraging characteristics (Svanbäck, Eklöv, 2002; Costa, Cataudella, 2007; Farré *et al.*, 2015). Foraging is strongly related to morphological traits (Mittelbach *et al.*, 1992), because they constrain feeding efficiency and the consumption of food items (Stoner, Livingston, 1984; Ehlinger, 1990; Pouilly *et al.*, 2003). Therefore, environmental characteristics, including habitat degradation, may influence the diet and morphology of species simultaneously.

To study these different factors that act upon the diet of fish species, it is important to have a study area containing several sampling points with different habitat characteristics

and degrees of environmental impact. A broad study area, such as the Campos Sulinos, is ideal for comparisons at these ecological scales. Furthermore, the characid *Bryconamericus iheringii* (Boulenger, 1887) is a good option for this study, since it occurs in a broad geographical area (Laguna dos Patos System, Uruguay River basin and Paraguay River Basin) and in different types of environments including small and big streams as well as shallow and deep lagoons (Casciotta *et al.*, 2002). *Bryconamericus iheringii* is considered a generalist, with opportunist dietary habits and a high degree of trophic plasticity (Oriccolli, Bennemann, 2006; Oosterom *et al.*, 2013), which are favorable characteristics for this study since they may reflect the availability of resources in the environment.

The objective of this study is to describe the diet and morphology of *B. iheringii* across a broad spatial scale to evaluate the influence of different environmental characteristics, such as substrate composition of the streams, percentage of vegetation cover, presence of human interference, etc. Thus, our hypothesis is that the diet and body shape of populations of *B. iheringii* will vary in response to different environmental characteristics across the Campos Sulinos ecosystem.

Material and Methods

Study area and sampling. According to Pillar, Lange (2015), the Campos Sulinos is a grassland ecosystem found on the three southernmost states of Brazil (Paraná, Santa Catarina and Rio Grande do Sul states), and distributed on two distinct biomes: the Atlantic Forest and Pampa. The grasslands from the Pampa biome extend throughout the Republic of Uruguay and Argentine provinces, from Bahia Blanca, Argentine to Porto Alegre – RS, Brazil, and is dominated by low grass and small shrubs.

For the present study, 22 streams belonging to two drainages (Tab. 1) and distributed in ten municipalities, in Rio Grande do Sul State, were sampled (Fig. 1). These streams are in the Pampa biome, which is distinguished by vegetation composed of grass, creepers, and some trees and bushes close to bodies of water (Pillar *et al.*, 2009). The climate of the area is subtropical, characterized by mild temperatures and rainfall well distributed throughout the year. Soil, in general, is fertile, being frequently used for agriculture (Pillar *et al.*, 2009).

Fish were sampled during October 2013 (in streams of Alegrete (AL), Quaraí (QR) and Soledade (SL) municipalities), November 2013 (in streams of Santo Antônio das Missões (SM) municipality), December 2013 (in streams of Santana da Boa Vista (SV) municipality), January 2014 (in streams of Jaguarão (JG) municipality), February 2014 (in streams of São Gabriel (SG) municipality), March 2014 (in streams of Santiago (ST) and São Pedro do Sul (SS) municipality), April 2014 (in streams of Lavras do Sul (LV) municipality) using backpack electrofishing. Each sampling took place within a 150 meter passage, after the streams' upstream and downstream were blocked with 0.5 mm nets, preventing fishes from entering and leaving the reach. After being eu-

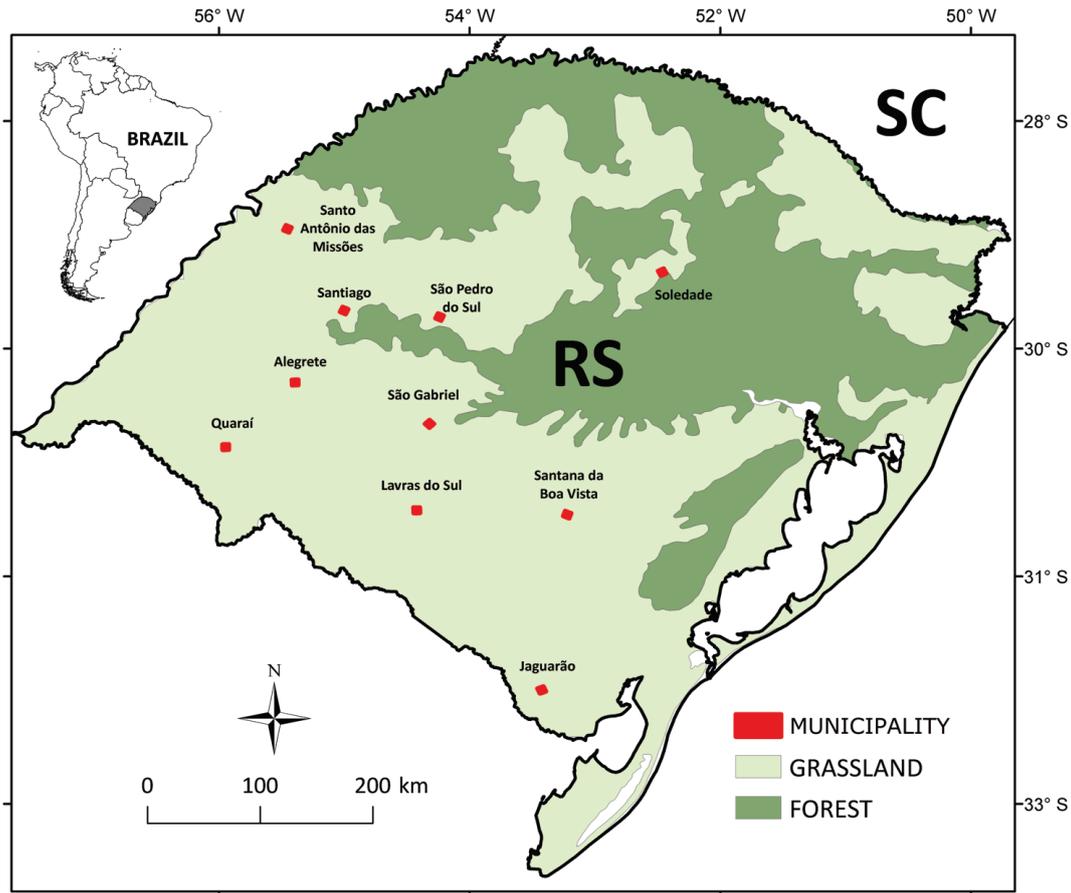


Fig. 1. Location of the ten municipalities in the Rio Grande do Sul State (RS), Brazil, and main land coverage. The landcover map was made with software ArcGIS 10.2 (ESRI, 2013) using the database of Pampa Biome cover of 2009 (UFRGS, 2016).

thanized with 10% eugenol (Vidal *et al.*, 2008; Lucena *et al.*, 2013), specimens were fixed in 10% formalin. At the laboratory, fish were preserved in 70% alcohol, identified with identification keys (Malabarba *et al.*, 2013; Serra *et al.*, 2014) and deposited in the fish collection of the Departamento de Zoologia at the Universidade Federal do Rio Grande do Sul (UFRGS 19238, UFRGS 19243, UFRGS 19246, UFRGS 19249, UFRGS 19254, UFRGS 19299, UFRGS 21339, UFRGS 21341, UFRGS 23695, UFRGS 21342). In addition, habitat patterns were measured, through previous geographical mapping and habitat characterization from inside the stream and its surroundings according to Kaufmann *et al.*, (1999) (see Dala-Corte *et al.*, 2016a) (Tab. 1, S1 and S2 – available only as online supplementary file accessed with the online version of the article at <http://www.scielo.br/ni>).

For the study of feeding characteristics, all adult individuals (standard length bigger than 45 mm) per stream were chosen (see Tab. 2 for abundance per stream). Only adults were analyzed in order to avoid ontogenetic influences. All individuals had their total weight (TW) (g) and standard length (SL) (mm) measured. Afterwards, all specimens were eviscerated.

Diet composition and analysis. Stomach contents were analyzed under optical and stereoscopic microscopes. For

invertebrates, identifications were made to the lowest taxonomic level possible, using taxonomic keys (Merritt, Cummins, 1996; Costa *et al.*, 2006; Mugnai *et al.*, 2010). Volumes of food items were quantified using graduated test tubes and a glass counting plate (Hellawell, Abel, 1971; Hyslop, 1980). Prey items were pooled into six categories: autochthonous invertebrates (Diptera larvae, Diptera pupae, Trichoptera, Ephemeroptera, Plecoptera, Odonata nymph, aquatic Coleoptera larvae, aquatic Lepidoptera larvae, aquatic adult Hymenoptera, Nematoda, testate amoebae, gastropoda, *Aegla*), autochthonous plants (diatomaceae algae, filamentous algae and aquatic plants), allochthonous invertebrates (adult terrestrial Hymenoptera, adult Diptera, adult Coleoptera, Araneae, etc), allochthonous plants (terrestrial plants), organic matter (scales, unidentified animal and vegetal matter, detritus (all organic matter decomposed)) and inorganic matter (sediment (sand or very small stones)) (Tab. 2).

In order to verify significant differences among streams, we used a Permutational Multivariate Analysis (PERMANOVA) with 9999 permutations (Anderson, 2001; Anderson *et al.*, 2008), based on a Bray-Curtis dissimilarity index (Anderson, 2001). This analysis was performed in R software (R CoreTeam, 2017) with the Vegan package (Oksanen *et al.*, 2009).

Tab. 1. Environmental variables measured for each stream in each municipality. %AGS: agriculture percentage in the upland catchment area; %MAR: percentage of forest surrounding the stream margins in a 50 meters wide buffer area; DEP: depth average of the stream (cm); WET: wetted width (cm); SBE: substrate cover percentage of boulders; SBO: substrate cover percentage of boulders; SCO: substrate cover percentage of cobble; SLP: substrate cover percentage of large pebbles; SSS: substrate cover percentage of small pebbles; SSA: substrate cover percentage of sand; SSI: substrate cover percentage of silt; SHA: substrate cover percentage of substrate hardpan; FCA: fish refuge composed of algae; FCM: fish refuge composed of macrophytes; FCW: fish refuge composed of wood; FCR: fish refuge composed of roots; FCL: fish refuge composed of leaves; FCH: fish refuge composed of hanging vegetation; FCU: fish refuge composed of a bank undercut; FLB: fish refuge composed of large boulders; FLO: velocity of water flow; CAO: openness of canopy above stream; LOI: percentage of anthropic impact surrounding the stream.

Stream	%AGS	%MAR	DEP	WET	SBE	SBO	SCO	SLP	SSP	SSA	SSI	SHA	FCA	FCM	FCW	FCR	FCL	FCH	FCU	FLB	FLO	CAO	LOI
AL01	0.32	45.32	48.42	3.15	48.38	11.55	3.86	2.28	4.11	6.25	0.63	17.67	0.00	0.00	11.36	23.41	20.23	37.95	11.59	20.91	24.91	0.06	20.00
JG01	2.47	43.37	20.91	3.16	4.06	0.00	1.00	17.51	15.94	52.61	0.42	0.00	0.00	0.00	8.64	23.64	11.59	10.68	13.41	4.55	68.00	0.04	5.00
JG02	3.51	67.03	15.06	4.27	0.00	0.00	0.00	0.86	46.99	45.18	0.49	0.00	0.00	0.00	15.45	18.41	8.64	21.36	5.00	0.00	33.73	0.07	10.00
LV01	42.96	23.35	31.28	5.30	0.00	0.00	0.00	0.44	25.71	63.42	0.44	0.00	0.00	0.00	34.55	20.68	10.00	5.45	0.45	0.00	70.00	0.05	110.00
LV02	6.05	57.26	41.21	5.25	1.93	1.66	50.33	31.22	3.92	9.52	0.76	0.00	0.00	13.86	10.45	30.23	0.00	42.73	7.73	5.00	104.91	0.23	5.00
LV03	5.17	77.93	38.82	6.66	0.00	0.00	0.00	13.18	20.64	61.01	0.00	0.00	0.00	0.00	26.14	27.27	2.73	28.41	12.05	0.00	54.82	0.23	20.00
QR01	20.45	12.59	34.06	5.16	1.62	1.88	14.52	36.24	24.05	0.00	4.27	0.00	0.00	12.73	28.64	28.41	17.73	24.77	10.00	0.45	15.73	0.14	112.50
QR03	47.57	43.77	20.09	3.45	25.42	10.93	49.19	14.46	0.00	0.00	0.00	0.00	0.00	0.45	1.82	6.82	7.73	33.18	10.23	33.86	16.27	0.35	232.50
SM02	4.01	66.94	29.12	3.20	0.00	0.00	0.00	0.49	30.41	30.40	10.09	0.00	0.00	0.00	45.68	29.55	44.32	14.09	23.86	0.00	0.00	0.04	0.00
ST03	5.13	45.84	30.00	3.91	8.49	0.00	0.00	29.30	12.27	10.40	0.00	0.00	0.00	0.00	15.68	26.14	3.18	11.59	9.77	0.91	28.82	0.04	120.00
SV01	3.42	41.75	8.06	3.57	0.49	0.00	2.17	38.25	25.23	33.27	0.00	0.00	0.00	0.00	6.82	1.82	0.91	2.73	0.45	0.45	31.27	0.66	80.00
SV02	0.61	80.74	14.15	4.15	0.00	0.00	0.00	0.00	7.80	80.82	0.00	0.00	0.00	0.00	30.45	18.41	15.91	15.91	11.82	0.00	18.00	0.15	55.00
SV03	5.24	64.53	17.88	4.72	15.40	1.70	10.24	21.50	28.32	18.20	0.44	0.00	0.00	0.00	8.64	21.59	18.18	15.45	12.27	1.36	43.73	0.26	10.00
SG01	2.32	18.40	4.08	0.89	0.00	0.00	0.00	0.00	0.00	68.58	25.87	0.00	0.00	0.00	30.91	11.82	4.09	23.64	0.00	0.00	8.00	0.14	137.50
SG02	5.57	75.34	7.85	2.09	12.70	0.00	0.47	5.88	5.41	59.81	6.34	8.89	0.00	3.18	15.23	2.27	3.64	3.64	0.00	5.23	6.91	0.24	250.00
SG03	2.33	57.00	7.32	1.55	0.00	0.00	0.00	4.40	9.15	76.14	4.41	0.00	0.00	0.00	8.18	14.55	13.41	13.64	12.27	0.00	0.27	0.03	240.00
SL01	45.30	66.12	43.61	6.42	12.10	7.45	54.19	20.39	3.92	1.95	0.00	0.00	0.00	0.00	19.77	11.59	1.82	15.23	22.73	5.00	72.27	0.19	10.00
SL02	37.40	47.65	32.39	6.67	1.60	0.70	44.47	26.12	9.67	17.43	0.00	0.00	0.00	0.00	23.41	18.86	5.00	18.86	18.64	2.27	31.82	0.03	277.50
SL03	32.19	31.73	26.73	4.01	7.27	1.39	37.36	36.49	6.51	10.97	0.00	0.00	0.00	0.00	24.55	24.77	7.27	25.45	36.82	3.18	18.82	0.05	65.00
SS01	15.85	26.90	16.00	6.15	0.00	35.78	43.68	16.11	2.30	0.00	0.00	0.00	0.00	0.00	4.09	8.41	6.36	8.64	2.27	17.73	40.09	0.10	10.00
SS02	30.29	62.28	39.26	7.79	0.00	3.03	9.43	10.56	13.65	28.49	24.61	0.61	0.00	0.00	19.55	22.50	1.36	6.82	21.36	0.91	7.55	0.04	0.00
SS03	56.66	43.36	22.43	4.20	5.92	43.92	36.15	3.01	3.01	2.33	0.00	0.00	0.00	0.00	4.55	8.18	2.27	4.55	17.95	17.73	22.64	0.03	0.00

Geometric morphometrics. We quantified the lateral body shape of *B. iheringii* populations among streams using landmark-based geometric morphometrics. For this procedure, the five largest individuals per stream were preserved in ethanol and photographed with a mounted camera (Canon PowerShot SX50 HS). We then used 16 homologous landmarks to quantify lateral body shape (Fig. 2). All analyses were performed in the tps program suite. Photographs were consolidated and landmarked using tpsUtil (Rohlf, 2004) and tpsDIG2 (Rohlf, 2006), respectively. Landmarks were superimposed, aligned and relative warps (RWs) were generated using tpsrelw (Rohlf, 2007). Relative warps are principal components of shape variation. Scale and rotation were removed from the analysis during superimposition and generation of the Procrustes fit.

Modelling. To simultaneously model and visualize the relationships between environmental, dietary, and morphological variables across the 22 streams, we used partial least squares canonical analysis (PLS-CA) (Rohlf, Corti, 2000). For this analysis, the relative proportions of prey items were pooled into the same generalized groups cited before: autochthonous invertebrates, allochthonous invertebrates, autochthonous plants, allochthonous plants, organic matter and inorganic matter. The environmental variables used to

modeling were in Tab. 1; first four principal components of body shape; and six major dietary categories. These were coded as blocks of variables and analyzed in a pairwise framework using the PLSCA function in the R package plsdepot (Sanchez, Sanchez, 2012).

Results

Adults diet composition. Diet of adults consisted of 28 food items. The most consumed groups were allochthonous plants (50.5%), followed by autochthonous invertebrates (25.2%) (Tab. 2). In the autochthonous invertebrates group, the most consumed items were aquatic insect remains, aquatic Lepidoptera larvae, aquatic Coleoptera larvae, Odonata nymphs and Diptera larvae (S3). Based on the PERMANOVA results, there were significant differences in diets among the 22 streams ($F = 22.709$, $p = 0.001$). The three streams in Soledade (SL) had a balanced contribution from organic matter and autochthonous items (both invertebrates and plants). The two Quaraí (QR) streams differed between themselves: QR01 had predominance of organic matter followed by higher volumes of autochthonous invertebrates, while QR03 had predominance of both autochthonous and allochthonous plants. The stream from Alegrete (AL), stream from Santo Antônio das Missões (SM) and the three streams

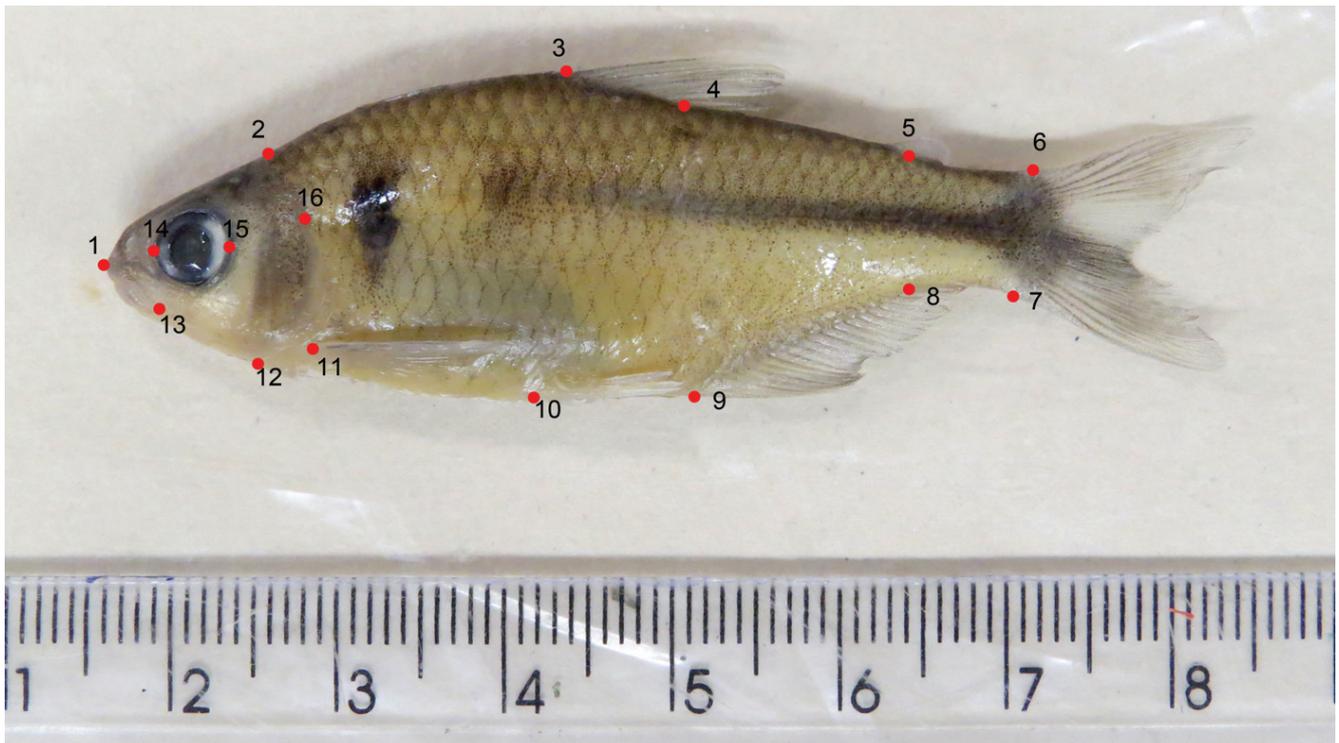


Fig. 2. Individual of *Bryconamericus iheringii* (UFRGS 21339) with the landmarks used for geometric morphometric. 1 = Tip of snout; 2 = Top of head (at the supraoccipital bone); 3 = Anterior insertion of dorsal fin; 4 = Posterior insertion of dorsal fin; 5 = Posterior insertion of adipose fin; 6 = Dorsal end of caudal peduncle; 7 = Ventral end of caudal peduncle; 8 = Posterior insertion of anal fin; 9 = Anterior insertion of anal fin; 10 = Origin of pelvic fin; 11 = Origin of pectoral fin; 12 = Ventral end of head; 13 = Tip of the upper jaw (maxillary bone); 14 = Anterior margin of eye; 15 = Posterior margin of eye; 16 = Anterior margin of opercle.

and canopy openness; however, was negatively correlated with the percentage of agriculture and substrate composed of cobble, boulder and bedrock (Fig. 4a).

Additionally, consumption of allochthonous invertebrates, organic matter, and inorganic matter were positively correlated with the presence of macrophytes, root cover, and substrate made of pebbles; and negatively correlated with local impact and the percentage of riparian forest (Fig. 4a). The consumption of autochthonous invertebrates was positively correlated with riparian forest and the presence of leaves (Fig. 4a).

Body depth and head length were positively correlated with substrate types such as silt and small pebbles, but negatively correlated with cobble, leaves and boulder cover, small boulders, bedrock and the percentage of agriculture (Fig. 4b). Pre-dorsal length was positively correlated with roots, hanging vegetation and woody cover; and negatively correlated with sand substrate, local impact, percentage of riparian forest and canopy openness (Fig. 4b). Mouth orientation was negatively correlated with roots, hanging vegetation and woody cover; however the positive correlations

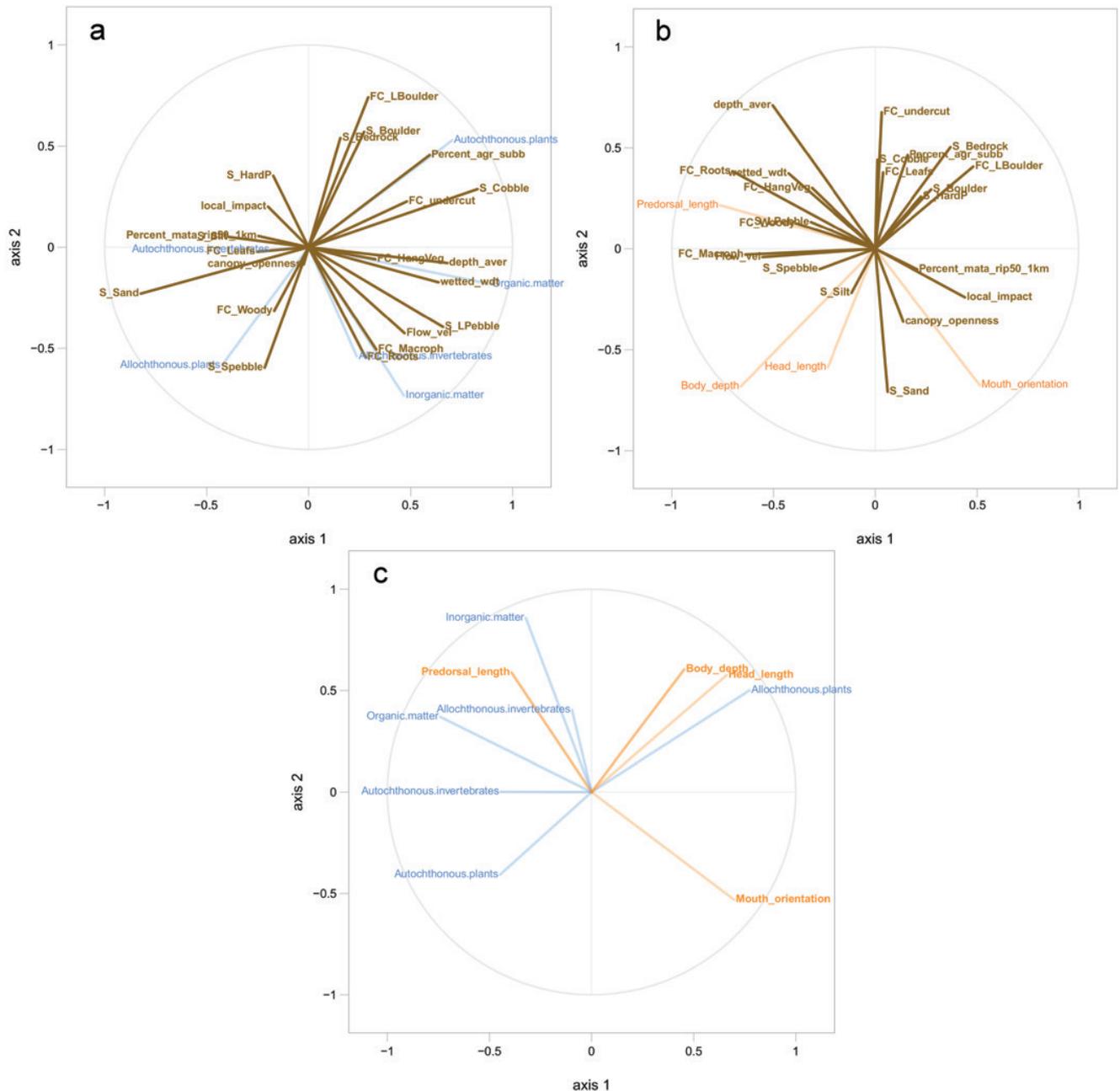


Fig. 4. Partial least squares canonical analysis depicting the covariation between: **a.** dietary (blue) and environmental (brown) variables; **b.** morphological (orange) and environmental (brown) variables; **c.** dietary (blue) and morphological variables (orange). Variables near the edge of the circle are better represented by the model. Variables near one another are positively correlated, whereas variables opposite one another are negatively correlated. Uncorrelated variables are orthogonal.

found for mouth orientation are possibly not biologically relevant (Fig. 4b).

Lastly, predorsal length was positively correlated with inorganic and organic matter and allochthonous invertebrates (Fig. 4c). Body depth and head length were positively correlated with the consumption of allochthonous plants and negatively with autochthonous plants (Fig. 4c). Mouth orientation was negatively correlated with the major of food groups.

Discussion

The diet of *B. iheringii* changed between streams that exhibited different environmental characteristics. However, body shape variation did not correspond to populations. In this way, our hypothesis that the diet and body shape of populations of *B. iheringii* will vary in response to different environmental characteristics across the Campos Sulinos ecosystem was not corroborated.

Out of all 22 streams sampled, in 14 (AL01, SM02, SV03, JAG01, JAG02, ST01, SS01, SS02, LV01, LV02, LV03, SG01, SG02, SG03) of them, *B. iheringii* had a high consumption of allochthonous plants with additional ingestion of autochthonous invertebrates and/or organic matter. All 14 streams have a high percentage of forest surrounding the margins in a 50 meters wide buffer area. In general, riparian forest, when well preserved, allows for more terrestrial resources to enter the streams; therefore, increasing the consumption of allochthonous items such as plants (Pusey, Arthington, 2003; Santos *et al.*, 2015). Moreover, these streams had null values for fish cover composed of algae and macrophytes, which may explain the lower consumption of autochthonous plants.

In the remaining streams (SL01, SL02, SL03, QR01, QR03, SV01, SV02, SS03), the diet of *B. iheringii* was based on high proportions of autochthonous invertebrates and plants with additional consumption of allochthonous plants and/or organic matter. These sites were also characterized by the presence of rocky substrate (boulders, cobbles or pebbles) and/or fish cover composed of vegetal matter (wood, roots and leaves). These conditions create a structured habitat where invertebrates' larvae can live and proliferate (Wallace *et al.*, 1997). Fallen leaves and other allochthonous items also create a structured habitat by falling inside streams, becoming food items and providing shelter and hideouts where invertebrates' larvae live, which consequently make them an available food item for *B. iheringii* (Allan *et al.*, 2003; Pusey, Arthington, 2003). Even though higher consumption of autochthonous plants is related with high canopy openness and loss of riparian vegetation (Gregory *et al.*, 1987; Allan, 2004), all streams from SL and stream QR01 presented intermediate values for canopy openness, low of riparian forest and hanging vegetation. It is possible that riverine primary production in all SL streams and in QR01 is higher than terrestrial primary production (Wootton 2012; Burrell *et al.*, 2014), making autochthonous plants a more viable foraging option for fish. Additionally, the two

streams of QR and streams SL01 and SL02 were the only sites where organic matter was as significant as the other food groups. Such organic matter could be from riparian forest (terrestrial origin), expressed as woods, leaves and roots as fish cover. It is also possible that in these streams the agriculture percentage is higher and that it contributes to environmental deviations in the aquatic environment, since more organic matter enters streams, which may increase foraging on food items such as detritus (Allan, 2004; Majdi *et al.*, 2015).

Such variation observed in the diet of populations of the same species in aquatic ecosystems is expected in environments with different compositions, as the Neotropical ones (Winemiller, 1990). This variety may be related to indirect interactions (Binning, Chapman, 2010), for instance, intraspecific variation related to morphology or variation in the structure of aquatic ecosystems (Oliveira *et al.*, 2010; Bonato *et al.*, 2017). Therefore, the results found are likely related to both these factors, since *B. iheringii* exhibit a variable diet that changes according to which food items are available in the environment (Oriccolli, Bennemann, 2006), as is usual for opportunistic fish (Rezende, Mazzoni, 2003; Ceneviva-Bastos *et al.*, 2010). Opportunistic and/or generalist fish species feed from a wide array of food items and they are also able to change their diet according to food availability (Karr, 1991; Pusey *et al.*, 2010). Accordingly, Rezende, Mazzoni (2003) found that even though *B. microcephalus* have an insectivorous diet, a few random food items were also consumed, reflecting the trophic plasticity of fish from the *Bryconamericus* genus in response to food availability in the environment. Nevertheless, *B. stramineus* have a diet specialized towards insects (Brandão-Gonçalves *et al.*, 2009). As previously discussed, our data showed that in environments with higher percentages of riparian vegetation or with substrate composed by leaf litter, where aquatic insects are found, *B. iheringii* tend to have an insect-based diet. Accordingly, in environments where macrophytes and algae were abundant, the species had a higher consumption of aquatic plants. Thereby, *B. iheringii* is considered a species that adapts to environmental alterations caused by anthropic impacts, since in the absence of items as insects, it forages on a high volume of detritus and algae (Dala-Corte *et al.*, 2016b).

PLS-CA analysis showed a relationship between diet and environmental variables. It was observed that feeding was related with substrate composition, percentage of forest surrounding the stream margin and composition of fish cover. For example, the percentage of agriculture surrounding streams was found to affect consumption of allochthonous and autochthonous plants. Accordingly, in streams with lower values of agriculture percentage in the upland catchment area, as the case of SL01, SL02 and SL03, there were small contributions of allochthonous plants as opposed to the high proportions of autochthonous plants, autochthonous invertebrates and organic matter. The ingestion of mainly autochthonous invertebrates was correlated with the percentage of riparian forest and, consequently, of fish cover composed

by leaves (which create a microhabitat for many aquatic insects). This pattern of feeding was observed in streams AL01, SM02, SV01, SV02, SV03 and SS03, where individuals of *B. iheringii* had the highest consumption of autochthonous insects and complemented their diets with allochthonous plants. Thus, the species exhibits different food items for different environmental characteristics. According to Casatti, Castro (2006), species in the *Bryconamericus* genus are considered active swimmers, thus allowing them to forage throughout the whole water column, but mainly from the substrate and from the surface, on items carried by the current (Ferreira *et al.*, 2012). Furthermore, this capacity to forage on a wide variety of items is also related to the species' morphology.

However, in this study we did not observe a structure of different body shapes among populations, likely due to comparatively high degrees of shape variation within populations. The two major feeding groups, formed by allochthonous plant and autochthonous invertebrate feeders, did not have different body shapes at the population-scale. There may have been minimal selective pressure for differences in body shape, as demonstrated by how similar body shapes were across sites. The food items observed in this study are found on the bottom of streams; thus, even though the predominant food group was not the same for all environments, the items were captured by *B. iheringii* from a similar habitat. Therefore, random phenotypic variation too different and/or far from the generalized shape may not offer high fitness in the majority of these habitats (Travis, 1994). Normally, fish with generalist diets have a more general body shape pattern, located between intermediate values for that specific environment (Faradonbe *et al.*, 2015). Intermediate body shapes differ from specialized body shapes associated with a specific habitat, which can decrease the efficiency of adaptation when habitats are heterogeneous or environmental conditions are unpredictable (Webb, 1984). Therefore, maintaining an intermediate, general body shape should allow for greater ecological flexibility for generalist fish that can utilize a wide array of prey (Wimberger, 1992; Tófoli *et al.*, 2013). Thus, such a generalized body shape probably offered advantages for foraging on a variety of items in streams with minor to large environmental differences.

Despite this low delimitation of specific shape among populations, the PLS-CA analysis showed that correlations between diet and morphology were associated with small variations in the body depth, body and head length, and mouth orientation. The specimens with a deep body and long head length tended to consume allochthonous plants and those with smaller values of those shape dimensions tended to consume items of autochthonous origin. Mouth orientation tends to be more subterminal due to the great consumption of items that are on the bottom, even portions of the allochthonous plants can be foraged from the bottom of the streams. This may be related to a greater efficiency in feeding on items in the middle of the water columns versus from the bottom (Neves, Monteiro, 2003; Pouilly *et al.*,

2003). Regarding mouth orientation, it tends to be more subterminal when there is a high consumption of items that are in the bottom (Loy *et al.*, 1998; Langerhans *et al.*, 2003), including allochthonous plants in banks of leaves on the bottom of streams. However, as already pointed, the effect of this relationship may be minimal. This pattern may arise due to the fact that feeding characteristics are labile such that they respond quickly to changes in the availability of prey items (Gerking, 1994). However, even if the species have a phenotypic plasticity that allows organisms to change their phenotype in response to shifts in the environment (Parsons *et al.*, 2016), these physical traits still respond more slowly than behavior, such as feeding. Nevertheless, other studies that include a majority of generalist and invertivore or omnivore species found the existence of a relatively weak relationship between diet and morphology (Douglas, Matthews, 1992; Motta *et al.*, 1995; Adite, Winemiller, 1997). In these studies, other factors may have had a major role in determining body shape and trophic guild of the species, such as their phylogenetic relationship or spatiotemporal variables. Therefore, opportunistic and generalist diets are frequently associated with a generalized pattern of body shape (Faradonbe *et al.*, 2015) as shown by the *B. iheringii*.

Moreover, many populations of generalists are composed by specialized individuals who feed on certain food groups belonging to the population niche (Araújo *et al.*, 2011). Intrapopulation variation can be caused by variation in resource use that seems to be intrinsic to the individual and has been described as individual specialization (Bolnick *et al.*, 2003). Such variation cannot be attributed to sex, age, environmental factors or how the individual uses the environment (Bolnick *et al.*, 2011). Individual variation can be elicited by several mechanisms such as genetic, phenotypic, environmental where the use of different resources by a single individual is constrained by functional morphology, cognitive ability or digestive ability (Svanbäck, Eklöv, 2003, 2004). The PLS-CA results showed a correlation between morphology and environmental variables, but these are likely weak patterns considering that populations largely overlapped in their body shapes. For example, in the primary body shape axis RW1 (body depth), the streams with more percentage of substrate composed by silt and small pebbles were expected to be in the positive scores. Nevertheless, in LV02 and SG01 there were low proportions of substrate composed by small pebbles and silt, but are located in the positive score area. For populations of *Eretmobycon emperador*, body depth also explained the largest variation among the samples (Robert, Hirtle-Lewis, 2012). The authors suppose that variations in body depth could be more related to regional environmental pressures than to other morphology characteristics that seem to vary among the population regardless to samples site. Other studies also found that the measured local habitat characteristics did not explain much about the observed morphological variation (Lostrom *et al.*, 2015), but are related to variation in other factors, such as interactions with environmental and ecological factors.

There was no strong correlation between diet and morphology with some environmental variables, such as percentage of agriculture and local impact. This is due to the fact that, despite being constantly threatened by soil exploitation, streams still have variable environments. Such variability can highlight the presence of a high proportion of riparian forest surrounding a wide area around the streams. It is known that riparian canopy and nearshore vegetation influence the trophic structure of stream fishes, because they may function as a resource and promote physical heterogeneity of the channel (Zeni, Casatti, 2014). This broad area of riparian forest is what most contributed to a large consumption of allochthonous plants by *B. iheringii*. However, if land exploration for livestock and soy production keeps rising, bigger morphological and dietary differences may become relevant (Allan, 2004; Dala-Corte *et al.*, 2016b).

The results found in this study indicate that morphology was not strongly correlated with environmental characteristics, while predominant food groups were correlated with only some environmental characteristics. Major axes of body shape (RWs) were correlated with some food items; however, the ecological relevance of such relationships may be limited since morphology seems to differ only slightly among populations. This low variability in morphology may be a reflection of the variety of resources available and the corresponding need to exploit numerous types of resources rather than specialize. An important question that should be addressed in future studies is the apparent individual variation within the populations of *B. iheringii*.

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